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Holocene mercury accumulation calibrated by peat decomposition in a peat sequence from the Sanjiang Plain, northeast China

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20 Abstract: Peatlands are ideal archives that are widely used to reconstruct historical Hg accumulation around the world. However, decomposition of peat soils 21 leads to Hg enrichment or depletion in peat profiles. To evaluate the impact of peat 22 decomposition on historical Hg accumulation records, a 7800-year peat sequence 23 from the Shenjiadian peatland (SJD-2, Sanjiang Plain, northeast China) was selected 24 in this study. Based on the degree of peat humification and a generalized additive 25 model (GAM), Hg accumulation rates (Hg ARs) were reconstructed and calibrated 26 from the middle Holocene onward. The results showed that the Hg concentrations in 27 the SJD-2 peat profile ranged from 11.9 to 55.3 ng g⁻¹ and that the Hg AR ranged 28 from 0.4 to 7.0 μ g m⁻² yr⁻¹; these values for both parameters were lower than their 29 corresponding values observed in other peatlands around the world. Peat 30 decomposition led to Hg depletion in peat soils to some extent, and the GAM could be 31 used to evaluate the impact of peat decomposition on historical Hg ARs in peat 32 sequences based on the degree of peat humification and calibration of Hg ARs. Before 33 the Industrial Revolution, anthropogenic Hg sources caused the calibrated Hg ARs in 34 35 the SJD-2 peat profile to slightly increase around 1300 cal. yr BP. Similar to other regions around the world, the calibrated Hg AR on the Sanjiang Plain also obviously 36 increased from 3 to 8 μ g m⁻² yr⁻¹ after global Hg emissions began to increase during 37 the Industrial Revolution. 38

Keywords: Mercury; Peat decomposition; Generalized additive model; NortheastChina

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42 **1. Introduction**

Mercury (Hg), which is an important trace element, was widely used in metallurgy and human activities several thousand years ago. Natural sources of Hg include volcanic activity, forest fires, natural weathering, and soil emissions, whereas anthropogenic sources of Hg include fossil-fuel combustion, industrial manufacturing, mining and smelting (Azoury et al., 2013; Pirrone et al., 2010). The raw material of mercury ores and other metal minerals present during ore smelting have caused large amounts of gaseous mercury to be released into the atmosphere (Mighall et al., 2002).

50 With the development of metal smelting, Hg produced by human activities continued to increase around the world from the Bronze Age to the Industrial Age (Drexler et al., 51 2016). Over large regions of the globe, human-related Hg emissions have increased 52 relative to natural sources since the Industrial Revolution, in the middle of the 18th 53 century (Elbaz-Poulichet et al., 2011; Mast et al., 2010). With the development of 54 industry and mining, increases in anthropogenic Hg production became the major 55 factor causing Hg accumulation rates (Hg ARs) in natural ecosystems to markedly 56 57 increase after the Industrial Revolution (Corella et al., 2017; Farmer et al., 2009). Gaseous elemental Hg can be maintained in the atmosphere for approximately 1 year 58 and can be widely transferred on a hemispheric scale (Martinez-Cortizas et al., 1999). 59 Because of the atmospheric circulation and long retention time of Hg in atmosphere, 60 large amounts of Hg produced from anthropogenic sources had exerted a serious 61 influence on Hg deposition globally, and high Hg ARs since the Industrial Revolution 62 have been widely reported in many regions around the world (Corella et al., 2017; 63 Daga et al., 2016; Elbaz-Poulichet et al., 2011). 64

65 In China, the earliest mining or smelting activities can be traced back to the Xia Dynasty in 2100 BC-1700 BC. The Iron Age began nearly 700 years later in 66 northeastern China than in the center of China, and the lifestyles of local residents 67 began to be influenced by Chinese culture from central China during the Qin Dynasty 68 (approximately 200 BC) (Xue et al., 1991). Historical Hg ARs increased in several 69 regions around the world when iron tools became widely used for human activities (Li 70 et al., 2016; Martinez Cortizas et al., 2016). However, few studies have focused on 71 historical Hg ARs in northeast China; thus, the relationship between regional human 72 73 activities and historical Hg accumulation in northeast China is still unclear. Historical variations in Hg ARs over the last several thousands of years can be observed using a 74 number of types of environmental archives (e.g., lake sediments, peatlands) (Drexler 75 et al., 2016; Mast et al., 2010). Similar to other ecosystems, peatlands are an 76 important environmental archive that are widely distributed around the world (Yu et 77 78 al., 2010). Due to the anaerobic environment of peatlands and their ability to receive dust via atmospheric deposition, trace elements are deposited and restored in 79

peatlands (Martini et al., 2007). Thus, peatlands have been widely used to reconstruct
the historical deposition fluxes of trace elements in previous studies (Bao et al., 2016;
Bao et al., 2012; Ohlson et al., 2006).

In addition to Hg deposition, previous works have suggested that Hg 83 concentrations in peatlands are also affected by the degree of peat decomposition 84 (Biester et al., 2003; Zaccone et al., 2009). It is also difficult for most age-depth 85 models to consider and modify the influence of peat decomposition on peat 86 accumulation rates (Blaauw and Christen, 2011). Thus, peat decomposition means 87 that the historical Hg ARs reconstructed based on peat accumulation rates and Hg 88 concentrations may not directly reflect historical Hg deposition. To decrease the 89 impact of peat decomposition on the results of Hg AR reconstruction, the C/N mass 90 ratio has been successfully used to evaluate the degree of peat decomposition and to 91 calibrate the reconstructed results of historical Hg ARs in previous studies (Biester et 92 al., 2003). However, the mass ratio of C/N cannot be employed to evaluate the degree 93 of peat decomposition based on the characteristics of organic matter and might affect 94 95 the calibrated Hg ARs. Similar to the mass ratio of C/N, the degree of peat humification is influenced by the characteristics of organic matter and is widely used 96 as another indicator to evaluate the degree of peat decomposition in peatlands (Biester 97 et al., 2014). Nevertheless, few studies have focused on the calibration of historical 98 99 Hg ARs through peat decomposition indicators other than the mass ratio of C/N.

In this study, we selected a ¹⁴C-dated profile (SJD-2) (Zhang et al., 2014) from the 100 Shenjiadian peatland (Sanjiang Plain, northeast China) for the determination of Hg 101 concentrations and estimation of historical Hg ARs over the past 7800 years on the 102 Sanjiang Plain, northeast China. The degree of peat humification and the mass ratio of 103 C/N were selected as two indicators for evaluation of the degree of peat 104 decomposition in the SJD-2 peat profile. Based on a generalized additive model 105 (GAM), historical Hg ARs were calibrated based on selected peat decomposition 106 indicators in this study. Combining the calibrated Hg ARs with regional potential 107 108 natural or anthropogenic Hg sources, the historical Hg sources and Hg ARs of the Sanjiang Plain (northeast China) were revealed. Finally, this paper provides some data 109

110 on Hg ARs and Hg concentrations in the peatlands of northeast China.

111

112 **2. Materials and methods**

113 **2.1. Site and sampling**

The SJD-2 peat profile was sampled in the Shenjiadian (N46°34'54"; 114 E130°39'44") peatland in Sanjiang Plain, Northeast China (Fig. 1). Detailed 115 information about the Shenjiadian peatlands can be found in Zhang et al. (2014). The 116 117 studied site is located in the valley zone between Zhangguangcai ridge and Wanda Mountain, which has an annual mean temperature of 2.4 °C and an annual mean 118 precipitation of 550 mm. In March 2010, the SJD-2 core was collected by peat 119 digging (Zhang et al., 2014). The profile was recovered to a depth of 195 cm and 120 sliced into sections that were 1 cm thick. This profile comprises silt sediments below 121 160 cm and peat above this depth. Samples were freeze-dried (at -40 °C and a 122 pressure of approximately 50 bar) for one week; they were then milled (using an agate 123 mortar and pestle) and homogenized. 124

125

126 **2.2. Hg analysis**

For Hg analysis, we measured all samples in the upper 10 cm, one of every two 127 samples between 11-90 cm, and one of every five samples below 90 cm. A total of 73 128 129 samples were analyzed in this study. Total Hg concentrations were determined in freeze-dried samples using a DMA-80 mercury analyzer (Milestone, Italy). One of 130 every ten samples was analyzed in duplicate as a control, and two standard reference 131 materials (M2 and M3) of the moss Pleurozium schreberi (Steinnes et al., 1997) were 132 run within each set of samples. The quantification limit was 0.01 ng g^{-1} and the mean 133 recoveries were 70% \pm 2.1% and 93% \pm 4.8%, respectively. The differences between 134 duplicates ranged from 4% to 9%, with a mean value of 6.3%. 135

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137 2.3. Bulk density, LOI, N, and degree of peat humification

To determine peat bulk density, peat cylinders were dried at 105 °C to a constant weight. The dry weight and known volume of the peat cylinders were used to

calculate their bulk density. Loss on ignition (LOI) was determined by burning 140 approximately 0.5 g of dry peat at 550 °C for 6 h, and the ash contents are the residual 141 materials left after LOI analysis. The carbon (C) concentrations were calculated from 142 LOI (i.e., multiplying the organic matter content by 0.5) (Lamarre et al., 2012). Total 143 nitrogen (TN) was determined through H₂SO₄ - K₂SO₄ - CuSO₄ - Se digestion in a 144 Continuous Flow Chemical Analyzer, and the mean recovery were 97 - 103%. The 145 degree of peat humification was estimated by performing colorimetry on peat alkaline 146 147 extracts (Chambers et al., 2011). Prior to peat humification analysis, the dry peat samples were milled to particle sizes that were smaller than 180 µm. A total of 0.125 g 148 of each peat sample was treated with 8% NaOH solution (50 ml, 90 °C) for 1 h. Each 149 mixture was filtered and transferred to a 100-ml flask that was then filled with pure 150 water. Then, the solutions were diluted with equally pure water and measured using a 151 UV-2500 visible spectrophotometer at a 540-nm wavelength. The absorbance values 152 of the solutions were characterized by their degree of peat humification (Vorren et al., 153 2012). 154

155

156 2.4. Chronology

To establish an age-depth model, eight samples were prepared following the 157 protocol described by Zhou et al. (Zhou et al., 2002), and they were measured by 158 acceleration mass spectrometry (AMS) in the Xi'an Accelerator Mass Spectrometry 159 laboratory (Zhang et al., 2014). Their radiocarbon ages were calibrated to calendar 160 ages using the CALIB Rev.7.0.1 software (Stuiver and Reimer, 2006) (Table 1). The 161 chronology of the sequence was calculated using the cubic-spline method based on 162 the 8 calibrated calendar age values, and the results of the age-depth model and peat 163 accumulation rates were shown in previous studies (Zhang et al., 2014; Zhang et al., 164 2017). 165

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167 2.5. Generalized additive model

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Generalized additive models (GAMs) are regression models where smoothing

splines are used instead of linear coefficients for covariates (Hastie & Tibshirani,
1990). This approach has been found to be particularly effective at handling complex,
non-linear behaviors associated with ecological and environmental research (Gao et
al., 2014b; Pearce et al., 2011). The additive model, in the context of a concentration
time series, can be written in the form (Hastie & Tibshirani, 1990):

$$\log(y_i) = \sum_{i=1}^n s_j(x_{ij})^2 + \varepsilon_i$$

where y_i is the Hg AR in the ith sample, $s_i(x_{ij})$ is the smoothing function of the ith 174 value of covariate j, n is the total number of covariates, and ε_i is the ith residual with 175 $var(\varepsilon_i) = \sigma^2$, which is assumed to be gamma distributed. Smoothing functions are 176 developed through a combination of model selection and automatic smoothing 177 parameter selection using penalized regression splines, which optimize the fit and 178 attempt to minimize the number of dimensions in the model (Wood, 2011). The choice 179 of the smoothing parameters is made through restricted maximum likelihood, and 180 95% confidence intervals are estimated using an unconditional Bayesian method 181 (Wood, 2011). This analysis was conducted using the 'gam' modeling function in the 182 R environment for statistical computing with the package 'mgcv' (Team, 2015). 183

184

185 **3. Results**

186 **3.1. Total nitrogen, mass ratio of C/N and degree of peat humification**

The total nitrogen (TN) content, mass ratio of C/N and degree of peat humification in the SJD-2 peat profile are shown in Fig. 2c-e. The concentrations of TN were low at the surface (0-12 cm) and bottom (171-205 cm) of the profile, which yielded TN concentrations ranging from 2.2-7.2 mg g⁻¹ and 0.8–7.1 mg g⁻¹, respectively. Between 13 and 170 cm, the TN concentrations fluctuated between 15 and 25 mg g⁻¹. The variations in LOI were similar to those in TN, and the LOI values fluctuated approximately 60% in the middle of the peat profile (5 - 165 cm). At the

bottom of the profile, LOI markedly decreased to close to 10%.

The carbon concentrations determined in this study were calculated from LOI. 195 Based on carbon concentrations and TN concentrations, the mass ratios of C/N were 196 calculated and are shown in Fig. 2e. In the SJD-2 peat profile, the mass ratio of C/N 197 ranged from 10 to 30, and low mass ratios of C/N were found in the upper layers. The 198 mass ratio of C/N was approximately 25 in the bottom layers (lower than 100 cm), 199 which was slightly higher than that in the upper layers, of approximately 20. The 200 degree of humification exhibited similar variation trends to the mass ratio of C/N. The 201 degree of humification was also low (i.e., lower than 15%) in the upper layers and 202 increased with increasing depth. The highest value (55.9%) was found at a depth of 203 97-98 cm (Fig. 2f); then, the degree of humification decreased with increasing depth. 204

205

3.2. Hg concentrations and accumulation rates

The average concentration of Hg was 25.1 ± 10.3 ng g⁻¹. The maximum 207 concentration existed in the surface layers (i.e., the ca. 3 cm layer; 55.3 ng g^{-1}), and 208 the minimum value appeared at ca. 50 cm (11.9 ng g^{-1}) (Fig. 3). Hg concentrations 209 sharply decreased in the surface peat layers, and the Hg concentrations in the deeper 210 peat layers (ca. 20 cm) were lower than 20 ng g⁻¹. As the depth of the peat layers 211 increased from 10 to 80 cm, the Hg concentration increased from 10 to 20 ng g^{-1} . 212 From 80 cm to the bottom, the Hg concentration showed slight fluctuations between 213 20 and 30 ng g^{-1} . 214

The peat accumulation rates in SJD-2 were calculated based on the age-depth 215 model and dry bulk density (Fig. 2a). The Hg AR ranged from 0.5 to 7.0 μ g m⁻² yr⁻¹, 216 with an average value of $2.2 \pm 1.4 \ \mu g \ m^{-2} \ yr^{-1}$. From the surface layers to 15 cm, the 217 Hg AR decreased from 7.0 to 2.0 μ g m⁻² yr⁻¹. Between 15 and 80 cm, the Hg ARs 218 were low (1.0 - 2.0 μ g m⁻² yr⁻¹) and exhibited no obvious fluctuations. An obvious 219 peak of the Hg AR appeared approximately 95 cm, and the highest Hg AR was 6.8 µg 220 $m^{-2} vr^{-1}$, which occurred at 94 cm. From 100 cm to the bottom, the Hg ARs fluctuated 221 and were slightly higher than those in the middle of the peat profile (15 - 80 cm). 222

224 4. Discussion

4.1. Mercury concentrations and accumulation rates

The range of mercury concentrations in the SJD-2 peat profile $(11.9 - 55.3 \text{ ng g}^{-1})$ 226 was lower than those reported for other peatlands in nearby regions or at similar 227 latitudes. For example, in the ombrotrophic peatland along the nearby Hongqi River 228 in the Small Hinggan Mountains in northeast China (west of the Sanjiang Plain), the 229 maximum recorded mercury concentration was 574.4 ng g⁻¹, and a major increase in 230 the mercury concentration occurred from 9 cm to the surface, with an average 231 mercury concentration of 111.3 ± 64.7 ng g⁻¹ in deeper soils (Tang et al., 2012). In the 232 middle of China (i.e., the Dajiuhu peatland), the Hg concentrations in surface peat 233 soils range from 100-595 ng g⁻¹ (Li et al., 2016) (Table 2), which is much higher than 234 those in the SJD-2 peat profile in northeast China. The Hg concentrations in the 235 SJD-2 peat profile were also lower than those in other peatlands around the world. In 236 the surface layers of peat soils that accumulated after 1750 AD in particular, Hg 237 concentrations are obviously increased around the world, to values of even higher 238 than 100 ng g^{-1} (Allan et al., 2013). Based on these Hg concentrations, the ecological 239 risk of Hg pollution in the SJD-2 peat profile is much lower in other peatlands around 240 the world. 241

The elements that accumulate at the bottoms of peat profiles can be regarded as 242 being influenced by few anthropogenic factors and can be used to determine the 243 historical accumulation rates of natural elements (Daga et al., 2016; Gao et al., 2014a). 244 In the SJD-2 peat profile, lower Hg ARs were found in the bottom layers of peat soils 245 that accumulated before 2000 cal. yr BP (ranging from 0.4 to 4.4 μ g m⁻² y⁻¹) (Table 2). 246 The low Hg ARs in the SJD-2 peat profile were similar to those reported from other 247 sites before 1500 AD, such as the Pyenean (Europe) and Belgium (Europe) sites, 248 where the Hg ARs were 1.5 - 2.5 and $1.3 - 2.4 \mu g m^{-2} v r^{-1}$, respectively (Enrico et al., 249 2017). A natural Hg AR of approximately 1 μ g m⁻² yr⁻¹ has been reported in peatland 250 in Greenland (Europe) from before 1750 AD (Pérez-Rodríguez et al., 2018), and the 251 Hg AR ranges from $1.0 - 1.4 \ \mu g \ m^{-2} \ vr^{-1}$ in Peru (South America) (Beal et al., 2014). 252 The natural Hg ARs measured in the SJD-2 peat profile were higher than those as 253

these sites. However, the natural Hg ARs in the Dajiuhu peatland (before 3000 cal. yr 254 BP), which is located in the middle of China, range from 0.5 to 12 μ g m⁻² yr⁻¹ and are 255 higher than those in the SJD-2 peat profile (Li et al., 2016). In addition, the Hg ARs in 256 the SJD-2 peat profile were also much lower than those observed in several sites in 257 France (Europe) and Patagonia (South America), where the Hg ARs were more than 258 10 μ g m⁻² yr⁻¹ before the Industrial Age (Daga et al., 2016; Elbaz-Poulichet et al., 259 2011). The one of the important potential factors is that the natural Hg sources (e.g., 260 261 forest fires, natural weathering) may major influenced by climate characteristics and the location of studied regions. The similar latitude and climate characteristics in 262 northeast China with other selected regions around the world cause the major Hg 263 natural sources in these regions are similar, while, obvious different with middle and 264 southern regions of China. Overall, the Hg ARs in northeast China were lower than 265 those in the middle and southern regions of China and were similar to the average Hg 266 ARs around the world when the major sources of Hg were natural sources. 267

Comparing the Hg ARs during different periods in the SJD-2 peat profile reveals 268 that the anthropogenic sources of Hg began to influence the Hg ARs on the Sanjiang 269 Plain around 2000 cal. yr BP, when the Iron Age started in northeast China (Tarasov et 270 al., 2006). After 2000 cal. yr BP, the Hg ARs on the Sanjiang Plain increased from 1.0 271 to 7.0 μ g m⁻² yr⁻¹. The Hg AR in the SJD-2 peat profile recorded during the Industrial 272 Revolution was approximately 7 μ g m⁻² yr⁻¹, and the maximum Hg AR occurred 273 around 100 cal. yr BP, which was far lower than those observed in other sites around 274 the world (Table 2). The maximum Hg AR in the SJD-2 peat profile occurred in the 275 1800s A.D., together with high global Hg production during the Industrial Age. 276 However, in recent years, the Hg AR decreased to 6.0 μ g m⁻² yr⁻¹ and exhibited trends 277 that were slightly different than the historical Hg AR variation trends observed at 278 other sites during this period. In South America, Hg ARs range from $3.4 - 5.0 \ \mu g \ m^{-2}$ 279 yr^{-1} in lake sediments in Peru corresponding to the time at which anthropogenic Hg 280 began to influence Hg accumulation (post-1800 A.D.); Hg ARs in Patagonia range 281 from 10 to 26 μ g m⁻² yr⁻¹ during this period (Hermanns & Biester, 2013; Daga et al., 282 2016). In addition to these two sites, the Hg ARs at most sites are higher than 10 µg 283

 $m^{-2} yr^{-1}$ or even higher than 100 µg $m^{-2} yr^{-1}$, which are both values that are obviously 284 higher than the Hg ARs in the SJD-2 peat profile from northeast China (Table 2). 285 Because there are more than 1000 kilometers between the Sanjiang Plain and the 286 middle of China, which is the cultural and industrial center of China, the development 287 of the economy and industry on the Sanjiang Plain occurred much more slowly than 288 in other regions in China (Li and Yuan, 1996). A few local anthropogenic Hg sources 289 around the Sanjiang Plain are responsible for most of the anthropogenic Hg in the 290 291 SJD-2 peat profile, while additional Hg from other regions arrives through atmospheric transmission. Thus, Hg ARs in the SJD-2 peat profile started to increase 292 together when global Hg production started to increase. Both natural and 293 anthropogenic Hg AR (especially anthropogenic Hg AR) in the SJD-2 peat profile 294 were lower than in other regions. The low Hg ARs and low Hg concentrations in the 295 SJD-2 peat profile cause the ecological risk of Hg pollution in the SJD-2 peat profile 296 to be much lower than those observed in other peatlands around the world. 297

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4.2. Geochemical characteristics of peat soils in the SJD-2 peat profile

To determine the basic geochemical characteristics and decomposition conditions of peat soils in the SJD-2 peat profile, bulk density, LOI, and TN were selected to reveal basic geochemical characteristics, while the mass ratio of C/N and the degree of peat humification were selected to evaluate decomposition conditions in this study. The scatter plot and correlation coefficients between these five indicators were plotted with the 'chart.Correlation' function in the R environment with the 'PerformanceAnalytics' package (Peterson and Carl, 2014) and are shown in Fig. 3.

Geochemical indicators, such as LOI/ash content, can be used to distinguish between minerotrophic and ombrotrophic peatlands. LOI in the SJD-2 peat profile ranged from 6% to 85%, with an average value of approximately 66%. The ash content is the residual material that remains after LOI analysis; the average ash content of these samples was approximately 34%. The high ash content in the peat clearly identified it as being minerotrophic because ombrotrophic peat usually contains much lower ash contents (5-20%) (Shotyk, 1996). Thus, the SJD-2 peatland

could be identified as a minerotrophic peatland. The bulk density values of most 314 samples were approximately 0.2 g/cm^3 . In addition, in a few bottom samples, the bulk 315 density ranged from 0.5 to 1.0 g/cm³ (Fig. 2). With an increasing bulk density, LOI, 316 TN and the degree of peat humification all decreased. In contrast to these three 317 indicators, the mass ratio of C/N decreased when the bulk density increased from 0.1 318 to 0.3 g/cm^3 , after which it increased markedly. LOI was found to be highly positively 319 correlated with TN and the degree of peat humification and negatively correlated with 320 321 bulk density (Fig. 3). However, there was no significant difference between LOI and the mass ratio of C/N. The mass ratio of C/N was only significantly different than TN, 322 and there was a weak negative correlation between TN and the mass ratio of C/N 323 (-0.32, P<0.01). Similar to LOI, the degree of peat humification was also highly 324 positively correlated with TN and negatively correlated with bulk density. 325

The degree of peat humification and the mass ratio of C/N are widely used to 326 evaluate peat decomposition in peat soils (Biester et al., 2014; Broder et al., 2012). 327 The mass ratio of C/N can also be used to evaluate carbon loss in peatlands, as a low 328 mass ratio of C/N reflects more carbon loss in a peatland (Biester et al., 2003). In the 329 SJD-2 peat profile, there is no significant difference between the degree of peat 330 humification and the mass ratio of C/N. When the mass ratio of C/N decreased from 331 27 to 22, the degree of peat humification increased from 40% to 45%. In addition, the 332 degree of peat humification fluctuated around 45% when the mass ratio of C/N 333 decreased from 22 to 18. With more carbon loss occurring in other peat layers, the 334 mass ratio of C/N was lower than 18, and the degree of peat humification also 335 decreased obviously. With an increasing degree of peat decomposition (reflecting 336 more carbon loss), the variation of the degree of peat humification (which first 337 increased and then decreased) indicated that the relationship between peat 338 decomposition and the degree of peat humification was not directly linear and was 339 similar to the relationship between Hg and peat decomposition. In Hg accumulation, 340 carbon loss in peat layers leads to more Hg accumulation in peat soils, and Hg ARs 341 342 can be calibrated by the mass ratio of C/N (Biester et al., 2003). Hg may also be lost with a high degree of carbon loss in peat layers, and the mass ratio of C/N may not be 343

suitable for calibrating Hg ARs in this type of peat sequence. A low degree of carbon loss causes the degree of peat humification to increase, and an opposite trend appears between a high degree of carbon loss and the degree of peat humification. Because of the similar trends of Hg and the degree of peat humification during peat decomposition, the degree of peat humification may act as a better indicator for calibrating Hg ARs in peat profile.

Because different climate characteristics lead to different degrees of peat 350 351 decomposition in peat sequences, several previous studies have reconstructed historical climate change based on the degree of peat humification, which can be 352 regarded as an indirect proxy (Vorren et al., 2012; Wang et al., 2010). Some results 353 have shown that a warm and wet climate may result in a low degree of peat 354 humification (Ma et al., 2009), whereas other studies have obtained opposite results 355 (Wang et al., 2010). Based on the relationship between the degree of peat 356 humification and peat decomposition previously discussed, we can hypothesize an 357 explanation for the opposite results obtained in previous studies. Warm and wet 358 359 climate characteristics could accelerate peat decomposition and increase the degree of peat humification; however, high peat decomposition can lead to more carbon loss 360 and decrease the degree of peat humification in peat profiles. The different stages of 361 peat decomposition may lead to different results between different peat profiles. 362 Because the degree of peat humification increased and then decreased with a decrease 363 in the mass ratio of C/N in the SJD-2 peat profile, the degree of carbon loss in the 364 SJD-2 peat profile could not be interpreted based on the degree of peat humification, 365 and the degree of peat humification could not be employed to reconstruct historical 366 regional climate characteristics directly. Thus, the degree of peat humification in the 367 SJD-2 peat profile could be only used to evaluate the degree of peat humification in 368 the peat profile and to calibrate Hg ARs in the peat profile in this study. 369

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4.3. Influence of peat decomposition on accumulated Hg

The ratio of C/N and the degree of humification are two indicators that are widely used to evaluate peat decomposition (Biester et al., 2014; Broder et al., 2012).

The above discussion indicates that there is a non-linear correlation between the mass 374 ratio of C/N and the degree of peat humification in the SJD-2 peat profile and that 375 these two indicators can be used to reflect the degree of peat decomposition in 376 peatlands under different conditions. Thus, both of these indicators were selected for 377 the GAMs. In addition, based on the Akaike information criterion (AIC) test, the final 378 models were obtained using the model based on these parameters showing the lowest 379 AIC value. The final model, which explained 69.3% of the observed deviance, can be 380 381 written as follows:

log(Hg AR) = s(time) + s(degree of humification)

No other refinements resulted in any significant improvement to the models, and 382 the results obtained with the GAM are shown in Fig. 4. Time (F=8.16, P<0.001) and 383 the degree of humification (F=3.94, P<0.05) were found to have significant effects on 384 the Hg AR. Similar to the hypothesis that the degree of peat humification is better 385 than mass ratio of C/N for calibrating Hg ARs, the GAM showed the same results and 386 divided the impact of peat decomposition on Hg ARs. The relationships of the Hg AR 387 388 with peat decomposition and accumulation time are shown in Fig. 4a. With an increasing degree of peat humification, the impact of peat decomposition on Hg ARs 389 was negative, and a high degree of peat humification, together with increased peat 390 decomposition, caused some Hg loss in peat soils and decreased the Hg AR. In 391 contrast, when the degree of peat humification increased from 30% to 50%, there was 392 no clearly decreasing trend of Hg, and the Hg ARs were not influenced by peat 393 decomposition when the degree of peat humification was lower than 30%. With an 394 increasing degree of peat decomposition, the negative impact on Hg ARs meant that 395 Hg was lost along with the total mass loss when the degree of humification increased 396 from 10% to 30%, which indicated that Hg concentrations in peatland may not be 397 influenced by peat decomposition. However, when the degree of peat humification 398 increased from 30% to 60%, the effect on total Hg was stable; thus, Hg concentrations 399 may increase with the total loss of peat mass. 400

401

To decrease the influence of peat decomposition on Hg AR, we predicted

historical Hg ARs with the same degree of humification using the GAM. Compared 402 with the original degree of peat humification, the degree of peat humification in the 403 prediction model was the same in all peat layers and ranged from 10% to 60%; the 404 results are shown in Fig. 5. When the degree of humification increased from 10% to 405 30%, Hg ARs decreased markedly. In contrast, there was no obvious decreasing trend 406 in Hg ARs when the degree of humification increased from 30% to 60% before 4000 407 cal. yr BP. From 4000 cal. yr BP to the present, a more obvious decreasing trend of 408 409 Hg ARs appeared as the predicted degree of humification increased from 10% to 30%, while no obvious decreasing trend of Hg ARs appeared when the degree of predicted 410 humification increased from 30% to 60%. During this period, the degree of 411 humification in the SJD-2 peat profile was higher than 30%. Similar to the previous 412 discussion, peat decomposition could not have caused the loss of Hg in these layers, 413 and the Hg ARs were mainly influenced by Hg deposition during this period. 414

In summary, increasing the degree of peat humification from 10% to 30% easily led to Hg loss in the peatland, and the GAM could be used to decrease the influence of peat decomposition on Hg ARs. Based on the GAM, the Hg ARs in peat profiles could be predicted and calibrated with the same degree of peat decomposition (i.e., the same degree of peat humification). Historical variations in the Hg ARs calibrated by the GAM with the same degree of peat decomposition could be regarded as the actual historical variations in Hg deposition fluxes.

422

423 **4.4. Potential sources of Hg in the SJD-2 peat profile**

Based on the GAM analysis, Hg is easily lost when the degree of peat 424 425 humification increases from 10% to 30% and remains nearly stable when the degree of peat humification increases from 30% to 60%. Considering the range of the degree 426 of peat humification in the SJD-2 peat profile, the Hg ARs predicted with the same 427 degree of peat humification (30%, boundary between two conditions) were used to 428 reflect regional historical deposition in this study. To evaluate the influence of 429 430 potential Hg sources on Hg ARs in the SJD-2 profile, historical population data from Heilongjiang Province (Li & Yuan, 1996), data on temperature (average chain length 431

of n-alkanes, ACL-ALK) and precipitation (proportion of aqueous biomarkers, Paq) 432 indicators reconstructed based on biomarkers on the Sanjiang Plain (Zhang et al., 433 2017), temperature data reconstructed from pollen in Jingbo Lake (west of the 434 Sanjiang Plain) (Li et al., 2011), wildfire frequency data reconstructed based on 435 charcoal on the Sanjiang Plain (Zhang et al., 2015), and East Asian summer monsoon 436 data reconstructed from stalagmite ¹⁸O values in the Dongge cave (Wang et al., 2005) 437 were used as indicators of historical climate and human activities, to evaluate the 438 439 influence of potential Hg sources on Hg deposition in the SJD-2 peat profile (Fig. 6).

The first peak values of calibrated Hg ARs appeared from 3800 to 2700 cal. yr BP 440 during the last 8000 years. Fires increase the concentrations of atmospheric 441 carbonaceous particulate matter and atmospheric mercury (Crutzen et al., 1990; 442 Lavous et al., 2000). Atmospheric mercury is influenced by fires and, thus, exhibits 443 good relationships with both the frequency and intensity of fires (Wang et al., 2010; 444 Cordeiro et al., 2002). Because intensive anthropogenic activities began to influence 445 the Sanjiang Plain around 1200 cal. yr BP (Cong et al., 2016), the frequency of fires 446 was mainly influenced by regional climate characteristics before 1200 cal. yr BP. 447 Based on lipid biomarkers and grain-size records in SJD-2, the climate on the 448 Sanjiang Plain was cold and wet from 6000 to 4000 cal. yr BP under the strong 449 summer monsoon, before shifting to warmer and drier conditions after 4000 cal. yr BP 450 451 (Zhang et al., 2017). The SJD-2 peat profile recorded a transformation from a shallow lake to peatland around 4000 cal. yr BP (Zhang et al., 2014; Zhang et al., 2017). 452 Several periods with high frequencies of fires recorded in the Honghe peatland on the 453 Sanjiang Plain occurred around 2700 and 800 cal. yr BP (Zhang et al., 2015). The 454 high frequency of natural fires from 3800 yr to 2700 cal. yr BP may be the major 455 reason that the Hg ARs increased to 2 μ g m⁻² yr⁻¹ during this period. 456

The amount of Hg produced by human activities during the Iron Age was obviously greater than that produced during the Bronze Age, which led the Hg ARs to increase around the world during the Iron Age (Martínez Cortizas et al., 2013; Martinez Cortizas et al., 2016). The Iron Age began around 2200 cal. yr BP in northeast China, nearly 700 years after it began in the middle of China and Europe

(Tarasov et al., 2006). After 2200 cal. yr BP, the population around the Sanjiang Plain 462 also started to increase (Li and Yuan, 1996), and the increasing human activities 463 caused more Hg to accumulate in the natural ecosystem. The Hg AR on the Sanjiang 464 Plain started to increase during the Iron Age, and an obvious peak appeared around 465 1300 cal. yr BP; the calibrated Hg AR increased to nearly 4 μ g m⁻² yr⁻¹ during this 466 period. The climate of the Sanjiang Plain was warm and dry around 1500 cal. yr BP, 467 which caused fires to occur more easily during this period (Li et al., 2011). One 468 469 famous war occurred on the Sanjiang Plain around 1288 yr BP, and the black carbon deposits and diagnostic ratios of polycyclic aromatic hydrocarbons (PAHs) in two 470 wetlands in the eastern region of the Sanjiang Plain were also shown to have been 471 obviously changed by regional human activities during this period (Cong et al., 2016). 472 The increasing human activities around the Sanjiang Plain and the widespread use of 473 iron tools in northeast China caused anthropogenic sources of Hg to increase after 474 2200 cal. yr BP. Thus, the increasing human activities and warm/dry climate 475 characteristics around 1300 cal. yr BP caused the calibrated Hg ARs to be higher than 476 those in surrounding periods (i.e., $4 \mu g m^{-2} v r^{-1}$). 477

From 400 cal. yr BP to the present, the frequency of fires was nearly stable and 478 slightly decreased, which means that the amount to Hg produced by natural fires was 479 nearly stable. However, the calibrated Hg AR increased from 3 to 8 μ g m⁻² yr⁻¹ during 480 this period. Similar to the increase in global Hg production (Allan et al., 2013; 481 Conaway et al., 2004), the increasing human activities in northeast China, together 482 with the rise in Hg production from anthropogenic sources, caused the calibrated Hg 483 AR in the SJD-2 peat profile to obviously increase. Furthermore, industrial factories 484 485 began to be established in northeast China around the 1900s AD (50 BP), and gold mining activities in the Great Hinggan Mountains, which are located in the western 486 region of the Sanjiang Plain, also started to produce gold in 1888 AD. Similar to other 487 sites around the world, the increased production of Hg from regional anthropogenic 488 sources caused the calibrated Hg AR to obviously increase over the past several 489 hundred years. Thus, anthropogenic sources of Hg were the major sources responsible 490 for increasing the calibrated Hg AR on the Sanjiang Plain from 400 cal. yr BP to the 491

492 present.

In summary, the high frequency of natural fires from 3800 to 2700 cal. yr BP 493 caused the historical calibrated Hg AR on the Sanjiang Plain to increase for the first 494 time during the last 8000 cal. yr BP. With the regional population increasing and iron 495 tools becoming more widely used by humans, human activities began to increase the 496 calibrated Hg AR on the Sanjiang Plain around 2200 cal. yr BP. Together with the 497 high frequency of natural fires, human activities caused an obvious peak period of the 498 499 calibrated Hg AR around 1300 cal. yr BP. With increasing human activities, the anthropogenic sources of Hg became the major sources of Hg and caused the 500 calibrated Hg AR to increase from 3 to 8 μ g m⁻² yr⁻¹ after 400 cal. yr BP. 501

502

503 **5. Conclusions**

In this study, the historical Hg ARs in the SJD-2 peat profile from the Sanjiang 504 Plain were reconstructed and calibrated based on the degree of peat humification 505 using a GAM. The Hg concentrations in the SJD-2 peat profile ranged from 11.9 to 506 55.3 ng g⁻¹, and the Hg ARs ranged from 0.4 to 7.0 μ g m⁻² yr⁻¹; the values of both of 507 these parameters were lower than their corresponding values in other peatlands 508 around the world. Peat decomposition could have caused Hg loss in peat soils and 509 influenced Hg accumulation. Based on the GAM, the Hg ARs modified by a unified 510 degree of peat humification were closer to historical Hg deposition and could, thus, be 511 used to reconstruct the Hg deposition history in a peatland. There were three periods 512 in which the Hg ARs on the Sanjiang Plain were higher than those in adjacent periods 513 (i.e., 3800 to 2700 cal. yr BP, around 1300 cal. yr BP and from 400 cal. yr BP to the 514 515 present). Similar to other sites around the world, anthropogenic sources of Hg mainly increased during the Iron Age and gradually became the major sources of Hg in the 516 most recent several hundred years. From 400 cal. yr BP to the present, the increase in 517 anthropogenic sources of Hg caused the calibrated Hg AR on the Sanjiang Plain to 518 increase from 3 to 8 μ g m⁻² yr⁻¹. 519

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Table 1. Radiocarbon dating by Accelerator Mass Spectrometry of peat core SJD-2. Depth shown from modern surface to bottom of peat sequence. Calibrated and conventional radiocarbon ages presented as years before present (yr BP, calibrated to 1950 A.D.). Calibrated dates indicated as median estimated age with 95% confidence limits. (Zhang et al., 2014).

	Depth(cm)	δ ¹³ C (‰)		pMC (%)		1	¹⁴ C Age (yr BP)		
Lab. Code		$\delta^{13}C$	Error (1σ)	рМС	Error (1σ)	¹⁴ C Age	Error (1σ)	cal yr BP	
XA7553	27	-37.78	0.51	93.39	0.28	550	24	541±20	
XA7592	60	-38.64	0.48	87.34	0.27	1088	24	1004±52	
XA7542	84	-30.36	0.37	84.2	0.27	1381	26	1308±31	
XA7543	107	-24.56	0.4	81.2	0.32	1673	32	1577±58	
XA7554	126	-22.74	0.56	81.16	0.24	1677	23	1577±49	
XA7555	145	-34.31	0.48	80.62	0.3	1731	30	1635±75	
XA7570	165	-30.14	0.36	64.34	0.21	3542	26	3857±46	
XA7571	195	-48.99	1.03	41.93	0.47	6982	90	7816±158	

Table 2. Hg concentrations and accumulation rates (AR) in peatland and lake sediments around the world. Based on the degree of local human activities, Hg was defined as natural Hg or anthropogenic Hg during different periods.

Location	Hg Concentrations	Hg AR	Hg AR Veer		Deferrere			
Location	ng/g	$\mu g/m^2.a$	rear	Waterials	Keierence			
Somiiona Dlain China Asia	14-31;	0.4-4.4;	8ka to 2ka BP;	Deat	This starlar			
Sanjiang Plain China, Asia	12-55	12-55 1.0-7.0 Post 2ka BP		Peat	This study			
Deiinhu China Asia	29-200;	0.5-12;	12ka to 3ka BP;	Deet	$(\mathbf{L}; \mathbf{s}; \mathbf{s} \mathbf{z})$			
Dajiunu China, Asia	100-595	5-36	Post 3ka BP	Peat	(LI et al., 2010)			
Down South America	31-32;	1.0-1.4;	12ka to 3.5ka BP;	Laka	(Pool at al 2014)			
Peru, South America	68-70	3.4-5.0	Post 1800 AD	Lake	(Deal et al., 2014)			
Datagonia South America	27-47;	8-15;	1600 to 1850 AD;	Laka	(Daga et al., 2016)			
Palagonia, South America	64-81	10-26	Post 1850 AD	Lake				
Datagonia South America	69-104;	30,	17ka to 5.4ka BP;	Laka	(Hermanns and			
Patagonia, South America	186	29	Post 5.4ka BP	Lake	Biester, 2013)			
California North America	50-150;	180;	Pre 1800 AD;	Tidal marsh	(Conaway et al., 2004)			
Camorina, North America	500-1200	560-10000	Post 1800 AD	Huai marsn				
Delaium Europe	35-46;	1.3-2.4;	Pre-1750AD;	Deat	(Allan et al., 2013)			
Beigiuni, Europe	29-1130	8.3-137	Post 1750AD	reat				
Spanish Europa	22-100;	20-80;	Pre 1300 AD;	Deat	(Martinez-Cortizas et			
Spanish, Europe	100-436	80-440	Post 1300 AD	reat	al., 1999)			
Duanaan Europa	38-218;	1.5-2.5;	10ka to 1500 AD;	Deat	(Enrice et al. 2017)			
Pyenean, Europe	50-238	6.0-40	Post 1750 AD	reat	(Effice et al., 2017)			
Creenland Europa	9-60	<1	Pre 1750 AD;	Deat	(Pérez-Rodríguez et			
Oreemand, Europe	47-297	1-9.3	Post 1750 AD	reat	al., 2018)			
Enon on Europe	20-129	3.3-115	Pre 1750 AD;	Lalva	(Elbaz-Poulichet et al.,			
Flance, Europe	36-577	41-462	Post 1750 AD	Lake	2011)			

1 Figure Captions

2 Figure 1. Location of Shenjiadian (SJD-2) peatland.

Figure 2. Variation of bulk density, LOI, total nitrogen, the mass ratio of C/N, the
degree of peat humification, Hg concentrations (Hg conc.), and Hg accumulation rates
in the SJD-2 peat core.

Figure 3. Correlation coefficients and scatter plots of dry density, LOI, total nitrogen
(TN), the mass ratio of C/N (C.N ratios), and the degree of peat humification
(Humification). On top, the absolute value of the correlation plus the results of the test
for an association between paired samples are indicated with stars.

Figure 4. Estimated smooth terms of Hg accumulation rates for the GAM. (a) The smooth function of calendar years. The y-axis represents historical deposition of Hg; (b) Smooth function of the degree of peat humification. The y-axis represents the impact of peat humification on Hg accumulation rates. The grey shadow is the 95% confidence interval.

Figure 5. Calibrated Hg accumulation rates based on peat decomposition proxies. a Historical calibrated Hg accumulation rates (ARs) with the same degree of peat humification (from 10 to 60%) predicted by the GAM; b historical variations in the degree of peat humification and the mass ratio of C/N in the SJD peat core; c historical variations in ash contents in the SJD-2 peat core.

20 Figure 6. Historical calibrated Hg accumulation rates in the SJD-2 peat profiles and their potential influencing factors. a Historical Hg accumulation rates (ARs) 21 calibrated by the GAM, with the degree of peat humification defined as 30%; b, warm 22 tree species in Jingbo Lakes, northeast China (Li et al., 2011); c, temperature (average 23 chain length of n-alkanes, ACL-ALK) and precipitation (proportion of aqueous 24 biomarkers, Paq) indicators in the SJD peatland reconstructed based on liquid 25 biomarkers (Zhang et al., 2017); d, Regional fire frequency recalculated based on 26 charcoal data in the Honghe peatland, north of the Sanjiang Plain (Zhang et al., 2015); 27 e, historical intensity of the East Asia monsoon reconstructed from DG $\delta^{18}O$ data 28 29 (Wang et al., 2005); f, historical population in Heilongjiang Province (summarized by Cong et al., 2016); g, Bronze Age, Iron Age, and Industrial Age in Europe, the center 30

31 of China and northeast China.











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41 Figure 3

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Figure 5



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50 Figure 6